

A Model for Optimal Treatment of Cassava Wastewater Using Anaerobic Baffled Reactor

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ABSTRACT: The major components of the effluents from cassava processing industries are cyanide and starch. However it is suspected that cyanide inhibits the treatment of cassava wastewater. The experimental data were successfully fitted to a polynomial model which was used to optimize the treatment processes at a laboratory scale. The Monod and Michealis-menten models for cassava wastewater treatment was successfully calibrated and validated in an ABR system. For Michealis-Menten model, the maximum substrate utilization rate is estimated in the range: 2866.88 to 1432.84 mg/l and for Monod's model, it is estimated in the range: 493 to 1242 mg/l, which is more realistic, hence validating the empirical model as more accurate than the former, which is theoretical. The result revealed that the inhibitor constant decreased from 9.9989 to 1.6101mg/l as the number of baffles increased from 3 to 10. To reach a maximum COD removal efficiency of 99%, it was found that the aspect ratio of 10, 20 baffles, cyanide inhibition constant of 30 mg/l and influent flow rate of 0.8 l/min, are the required optimum operating conditions of the anaerobic baffled reactors.

KEYWORDS: Anaerobic baffled reactor, Cassava Wastewater, COD removal, Optimal treatment, Kinetic coefficients

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Notation

S_o = Influent COD at the start of each of the experimental case, mg/l, X_o = Influent suspended solids at the start of each case, mg/l, I = Inhibitor (cyanide) concentration at the start of each case, mg/l, S = Effluent COD after the first step of each case, mg/l, X = Effluent suspended solids, mg/l, k_s = Half saturation constant, mg l^{-1} , μ = Specific growth rate of organism, day^{-1} , μ_{\max} = Maximum specific growth rate of organism, day^{-1} , k_d = Endogenous decay coefficient, minute^{-1} , Y = Cell yield coefficient, ratio of the mass of cells formed to the mass of substrate consumed, mg/mg, k = Maximum rate of substrate utilization per unit mass of micro-organisms, minute^{-1} , k_s = Half velocity constant, substrate concentration at one-half of the maximum growth rate, mg/l, K_r = Overall COD removal rate constant, minute^{-1} , K_m = Michaelis-Menten constant, mg/l, K_i = Inhibition constant, mg/l, V_r = Reactor volume, l, L/W = Length-width ratio, N_b = Number of baffles, Q = Flow rate, l/min, HRT = Hydraulic retention time, $\text{min} = V_w/Q$.

I. INTRODUCTION

In most developing countries of sub-Saharan Africa as well as Asian countries, cassava (*Manihot esculenta*) has become a popular source of Carbohydrates (Eze and Azubuike, 2010). Cassava wastewater is an industrial residue obtained during the processing of cassava into various fermented products such as Garri, Fufu. Ubalua (2007) suggested the use of microorganisms to convert cassava wastes to value-added components such as Methane (biogas), Ethanol, surfactants

and fertilizers etc. Adewoye et al. (2005) showed that the physio-chemical characteristics, including dissolved Oxygen ($\text{DO} = 0.74 \text{ mg/l}$), biochemical oxygen demand ($\text{BOD} = 185 \text{ mg/l}$), total suspended solids ($\text{TSS} = 87 \text{ mg/l}$), of the cassava wastewater deviated from the FEPA (1991) standard ($\text{DO} = 5.0 \text{ mg/l}$, $\text{BOD} = 50 \text{ mg/l}$, $\text{TSS} = 30 \text{ mg/l}$) for the maximum limit allowable for effluent discharge into water bodies.

Even Okunade and Adakalu (2013) asserted that cyanide has been always present in cassava wastewater as Cyanogenic Glycosides which hydrolyzes to harmful form as Hydrocyanic acid (Olayinka, 2013). In addition, the degraded cassava peels were said to contain hydrocyanic acid and generated objectionable odor (Okunade and Adakalu, 2013). According to WHO (2006), NIS (2007), high concentration of cyanide ($>0.01 \text{ mg/l}$) in water is very dangerous to human health because cyanide affects the Thyroid and the central nervous system; leading to paralysis. Oghenejoboh (2015) proved that cassava wastewater contained pollutants that are deleterious to natural fish population. High biochemical oxygen demand present in cassava wastewater, depletes the dissolved oxygen (DO) needed for the survival of aquatic life in the receiving water bodies (Adegoke et al., 2020).

However, Igwe and Azorji (2018) showed that the cyanide content of the cassava effluent could serve as an efficient source of nutrient to the soil and thus to crops, making it an alternative to mineral fertilizer. According to Ogbhodo et al. (2001), cassava effluent increased the number of organisms in the soil ecosystem as well as the soil pH, organic carbon and total Nitrogen, even favorable effects such as

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Table 2: Definition of the cases of cassava wastewater treatment.

| Reactor | Q | HRT | S _o | X _o | I | S | X |
|---------|--------|----------|----------------|----------------|-----|-----|---------|
| 1 | 0.4 | 25 | 2189 | 10680 | 120 | 408 | 412.896 |
| 2 | 0.3148 | 31.7662 | 2060 | 10560 | 109 | 384 | 388.608 |
| 3 | 0.2796 | 35.76538 | 2000 | 10503 | 98 | 368 | 372.416 |
| 4 | 0.2444 | 40.91653 | 1970 | 10467 | 88 | 352 | 356.224 |
| 5 | 0.2092 | 47.80115 | 1930 | 10410 | 76 | 332 | 335.984 |
| 6 | 0.1739 | 57.50431 | 1900 | 10032 | 64 | 308 | 311.696 |
| 7 | 0.1287 | 77.70008 | 1850 | 9600 | 52 | 300 | 303.6 |
| 8 | 0.1035 | 96.61836 | 1810 | 9207 | 43 | 276 | 279.312 |
| 9 | 0.0683 | 146.4129 | 1750 | 8796 | 31 | 240 | 242.88 |
| 10 | 0.0331 | 302.1148 | 1700 | 8400 | 20 | 204 | 206.448 |

By adopting the approach by Onukwugha (2015), the COD removal efficiency (RE) was formulated in Equation 1 as a polynomial function of ten-level experimental runs of four independent factors: the reactors' aspect ratio (L/W), number of baffles (N_b), the inhibition constant (k_i) and flow rate (Q).

$$RE = m(L/W)^{a_1} (N_b)^{a_2} (K_i)^{a_3} (Q)^{a_4} \quad (1)$$

According to Ibeje et al. (2019), the COD removal efficiency of the reactor was estimated for each of the 10 cases as:

$$RE = (S_o - S) / S_o \quad (2)$$

where the terms retain their meanings as given in the notation section.

According to Potivichayanon et al. (2020), the mathematical model for cassava wastewater treatment without considering inhibition is given as the Monod model as:

$$r_A = \frac{ds}{dt} = \frac{Q}{V} (S_o - S) = \mu \cdot X = \frac{\mu_{max} S}{K_s + S} X \quad (3)$$

where the terms retain their meanings as given in the notation section.

If K_s is replaced with K_m, Eq. (3) gives the Michaelis-Menten model and simplifying the expression gives:

$$\frac{xV}{Q(S_o - S)} = \frac{K_m}{\mu_{max}} \frac{1}{S} + \frac{1}{\mu_{max}} \quad (4)$$

The mathematical modelling of cassava wastewater treatment with cyanide inhibition is also given by Potivichayanon et al. (2020) as in Eq. (5).

$$\mu_g = \frac{\mu_{max}}{1 + \frac{K_s + I}{S_o} + \frac{I}{K_i}} \quad (5)$$

Replacing μ_g with μ in Eq (5) and substituting μ with its equivalent definition as given in Eq. (3), then the inverse and simplified form of Eq (5) is:

Substituting and replacing μ and μ_g, then taking inverse and simplifying the expression gives:

$$\frac{xV}{Q(S_o - S)} = \frac{1}{\mu_{max}} + \frac{K_s}{\mu_{max}} \cdot \frac{1}{S} + \frac{1}{\mu_{max} K_i} \quad (6)$$

From Tables 2 and 4, the intercepts and slopes of the graphs of Q(S_o-S)/V_w X_o against 1/HRT were used to obtain k_d and Y for each of the cases, using: K_d = intercept/slope and Y = 1/slope (Potivichayanon et al., 2020). To obtain k and k_s using: k = 1/intercept and k_s = slope/intercept, the intercepts and slopes of the graphs of HRT/(1+k_d*HRT) against 1/S were used (Potivichayanon et al., 2020). The slopes of the graphs of Q(S_o-S)/V_w against S were used to obtain K_r for each of the cases using: K_r = slope. Using: K_m = slope/intercept, the intercepts and slopes of the graphs of 1/Q against 1/S were used to obtain K_m. It should be noted that the variables were as defined in the notation section.

III. RESULTS AND DISCUSSION

A. Cassava Wastewater Characterization

Table 3 shows the results of the tests at the Imo State Polytechnic Umuagwo in 2019 on the raw wastewater prior to treatment. The COD and BOD concentrations in the wastewater are evidently very high, thus the free discharge of untreated wastewater would definitely lead to environmental pollution.

B. Results of Estimated Kinetic Coefficients

Generally, the values of kinetic constants presented in Tables 4 are within the normal range of ABR process found in previous studies by Ibeje and Okoro (2013). For Michealis-Menten model, the maximum substrate utilization rate is estimated in the range: 2866.88 to 1432.84 mg l⁻¹ at a coefficient of determination, R²% of 53% and for Monod's model, it is estimated in the range: 493 to 1242 mg l⁻¹ at R²% of 84%.

Table 4: Estimated kinetic coefficients of abr-treated cassava wastewater.

| Reactor | K _d | Y | k | k _s | K _m | K _r |
|---------|----------------|-------|------|----------------|----------------|----------------|
| 1 | 0.138 | 1.026 | 2.03 | 1243 | 2865.88 | 13.577 |
| 2 | 0.137 | 1.36 | 2.02 | 1171 | 2697.14 | 10.683 |
| 3 | 0.136 | 1.55 | 1.99 | 1111 | 2584.41 | 9.646 |
| 4 | 0.137 | 1.779 | 1.95 | 1040 | 2472.5 | 8.747 |
| 5 | 0.137 | 2.076 | 1.9 | 953 | 2331.72 | 7.849 |
| 6 | 0.138 | 2.414 | 1.85 | 853 | 2163.43 | 7.018 |
| 7 | 0.137 | 3.175 | 1.83 | 831 | 2107.1 | 5.192 |
| 8 | 0.138 | 3.828 | 1.79 | 740 | 1938.53 | 4.499 |
| 9 | 0.139 | 5.624 | 1.73 | 615 | 1685.74 | 3.369 |
| 10 | 0.142 | 11.19 | 1.66 | 493 | 1432.84 | 1.908 |

Table 3: Average characteristics of the cassava wastewater used in the study.

| Wastewater (m ₃ /l starch) | pH | Total solids (mg/1) | Suspended solids (mg/1) | Total dissolved solids (mg/1) | BOD (mgO ₂ /1) | COD (mgO ₂ /1) | Total nitrogen (mg N/1) |
|---------------------------------------|-----|---------------------|-------------------------|-------------------------------|---------------------------|---------------------------|-------------------------|
| 25 | 4.5 | 7604 | 2642 | 6483 | 3608 | 8842 | 172 |

Based on the previous studies in describing the treatment kinetics of the cassava wastewater by Ibeje and Okoro (2013), the range of the maximum substrate utilization rate obtained from the Monod's model appears to be more realistic than the range of values obtained from the Michealis-Menten model, hence validating the accuracy of the empirical model when compared with the former, which is theoretical. Figures 2 and 3 show the variation of the kinetic coefficient with the inhibition coefficient, K_i . The K_d and Y intersect in Figure 2 at a value of 2000 min^{-1} . The maximum growth rate was estimated to be between 0.138 to 0.142 min^{-1} for cassava wastewater treatment (Agunwamba et al., 2001). This value is a little lower than the values obtained by Onukwugha et al. (2016). Also, Mantzaris (2002), studied the effect of cyanide concentration on COD removal efficiency and obtained similar graphs as the one shown in Figure 4.

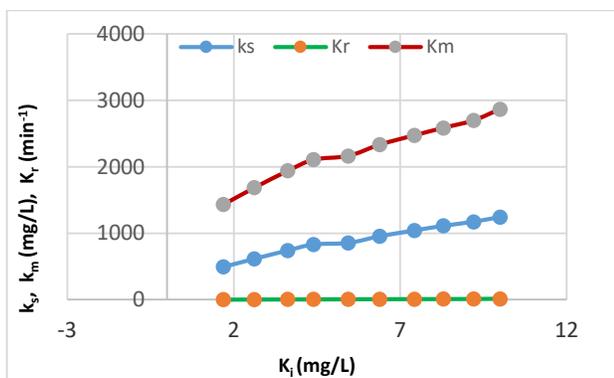


Figure 2: Influence of cyanide inhibition (K_i) on the rates of substrate consumption (K_s, K_m).

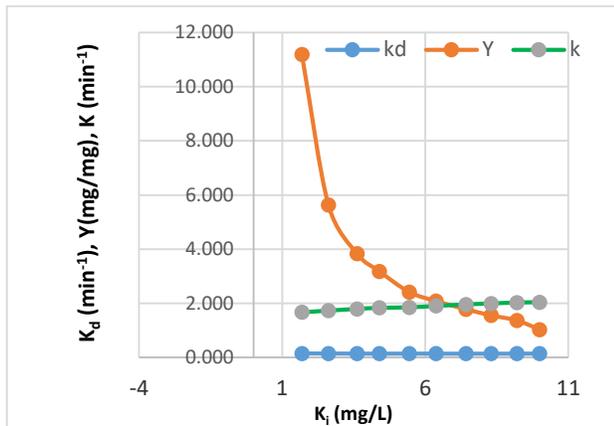


Figure 3: Cyanide inhibition (K_i) of cannibalism by microbes (k_d) and their rate of substrate uptake (K, Y).

According to Figure 4, at first glance, the point of deflection clearly indicates the occurrence of inhibition. Here the COD removal efficiency dropped from 88% to 84% between a cyanide concentration of 10 to 50mg/l. It remained constant between 50 and 70mg/l of cyanide until it further dropped to 82%, at which there was a gradual decline in COD removal efficiency.

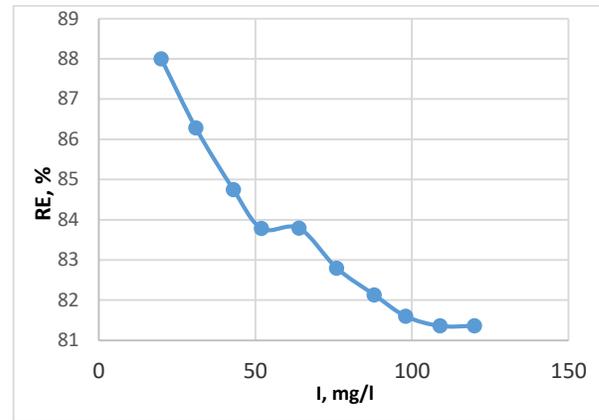


Figure 4: Effect of cyanide inhibitor on COD removal efficiency.

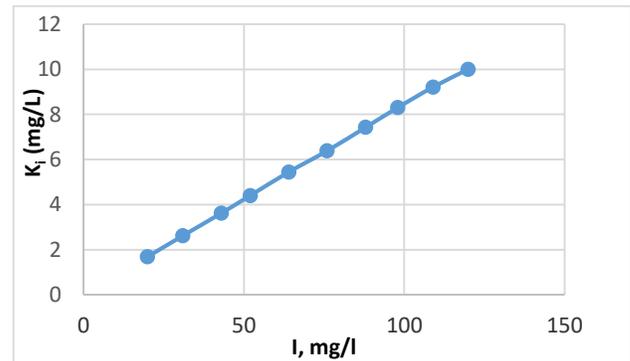


Figure 5: Degree of dependence of inhibitor constant, K_i on cyanide concentration, I .

C. Optimal Model for COD Removal in Cassava Wastewater

This was observed at cyanide concentration of 110mg/l. When Potivichayanon et al. (2020) compared the fixed-film sequencing batch reactor (F-SBR) with that of a conventional sequencing batch reactor (SBR), the cyanide and COD removal efficiencies of the SBR were 42.61% and 36.83%, respectively, while those of the F-SBR were 77.95% and 74.43% and the F-SBR was very effective for the complete removal of cyanide when the hydraulic retention time was increased to 10 days. This efficiency was similar to the efficiency of COD removal in this study, which reached 88–84% efficiency with the ABR system. These results showed that the immobilization of cyanide-degrading bacteria such as *Agrobacterium tumefaciens* SUTS 1 and *Pseudomonas monteilii* SUTS 2 carried out with a polypropylene ring in a fixed-film aerobic system enhanced the improved performance of the F-SBR.

Table 5 shows the four-factor, ten level experiment runs with observed values for COD removal efficiency. The data in Table 5 was used to calibrate the model that would govern the effect of the four parameters $L/W, N_b, K_i$ and Q on the COD removal efficiency, RE, as:

$$RE = 0.8304(L/W)^{0.0193} (N_b)^{0.0011} (K_i)^{0.012} (Q)^{0.016} \quad (7)$$

For the COD removal efficiency in cassava wastewater, positive values of 0.0193 and 0.0011 were respectively

estimated as the exponents of the aspect ratio (L/W) and number of baffles (N_b), implying that the COD removal efficiency increases exponentially as the number of baffles and aspect ratio of the reactor increases.

Table 5: 4¹⁰ Factorial model with observed cod removal efficiency.

| Case | L/W | N _b | K _i | Q | RE |
|------|-----|----------------|----------------|--------|----------|
| 1 | 0.4 | 3 | 9.9989 | 0.4 | 0.813614 |
| 2 | 0.5 | 4 | 9.2079 | 0.3148 | 0.813592 |
| 3 | 0.6 | 5 | 8.3002 | 0.2796 | 0.816 |
| 4 | 0.7 | 6 | 7.4252 | 0.2444 | 0.82132 |
| 5 | 0.8 | 7 | 6.3827 | 0.2092 | 0.827979 |
| 6 | 0.9 | 8 | 5.4414 | 0.1739 | 0.837895 |
| 7 | 1 | 9 | 4.3996 | 0.1287 | 0.837838 |
| 8 | 1.1 | 10 | 3.6196 | 0.1035 | 0.847514 |
| 9 | 1.2 | 11 | 2.6198 | 0.0683 | 0.862857 |
| 10 | 1.3 | 12 | 1.6901 | 0.0331 | 0.88 |

Thus, there is clearly a positive correlation between the COD removal efficiency in cassava wastewater, the aspect ratio (L/W) and number of baffles (N_b), although the aspect ratio shows higher correlation than the number of baffles. This agrees with the findings of Watthier et al. (2019) who used horizontal anaerobic fixed bed reactor to remove organic matter from cassava wastewater and the results showed that the COD removal efficiency increased with the organic loading rate, thus resulting in COD removal values of up to 99%. Conversely, cyanide inhibition constant (k_i) and the flow rate of the reactor inversely correlated with the COD removal efficiency (CE). These factors, thus have negative effects on the COD removal efficiency such that any decrease in any of the factors caused a decrease in the amount of COD removed.

The goodness of fit of the developed model was validated by estimating the coefficient of determination (R²) as 0.997 implying an adequate model. With the objective function for COD removal defined as Equation 8 and the range of independent functions given in Table 5 as the model constants, the response surface methodology was used to obtain the optimum condition for maximum COD removal using the numerical optimization tool built in the Design-Expert 9.0.4.1 statistical software. This was used to combine the desirable ranges for the response to obtain a simultaneous objective function as given by Myers et al. (2004) in Equation (8).

$$D = (d_1 \times d_2 \times \dots \times d_n)^{1/n} = (\prod_{i=1}^n d_i)^{1/n} \tag{8}$$

where *D*, *d_i*, and *n* are the desirability objective function, response range, and the number of responses, respectively.

In this case, the response which is the percent removal of COD (*d*₁) was maximized. Hence, the following optimum conditions to reach a maximum COD removal efficiency of 99% were found: reactor aspect ratio (10), number of baffles (20), cyanide inhibition constant (30 mg/l) and influent flow rate (0.8 l/min). Several authors have studied the effect of aspect ratio (L/w) on COD removal in ABR system and reported that the optimum aspect ratio for COD removal is the range of 0.5 to 1.5 (Marquardt, 1995; Parker and Doyle III, 2001; Fredrickson et al., 1970). Zamamiri (2002) studied the influence of number of baffles on COD removal efficiency in bioreactors and obtained the range of value 2 and 15. This is in agreement with the results obtained in this study.

IV. CONCLUSION

The objective of this study was the evaluation of the performance of the ABR during various hydraulic and loading conditions of cassava wastewater. A new model for optimal COD removal from cassava wastewater treatment was developed using for the ABR system using the RSM. Evaluation of optimum number of baffles and aspect ratio for the treatment of cassava wastewater was achieved,

The mathematical model for cassava wastewater treatment was successfully calibrated and validated in an ABR system. When fitted to experimental data, the Monod models showed a better description of cassava wastewater treatment in the ABR system than the Michealis-Menten model. The effect of N_b and L/W on the COD removal efficiency showed that their values increased as COD removal efficiency. Although, the impact of flow rate Q and K_i on the cyanide concentration revealed optimum values, further studies are required for the optimization of methane production from cassava wastewater treated in the ABR.

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